## Removal of Iron from Dyffryn Adda, Parys Mountain, N. Wales, UK using Sono-electrochemistry (Electrolysis with assisted Power Ultrasound)

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## Abstract

The Dyffryn Adda from Parys Mountain, N. Wales is one of the most polluting mine waters in the UK releasing c. 10 tonnes of copper per annum and 24 tonnes of zinc per annum into the Irish Sea. The Metal Mines Strategy for Wales has ranked it first. An acid, iron rich mine water visible by its ochreous staining along 3 km of the Afon Goch Amlwch to its coastal discharge at Porth Offeiriad (Priest Port) has a negative impact on both river and coastal water quality and local businesses and communities. Several investigations using Active, Passive and Hybrid treatment processes employing conventional treatment technologies as well as Pump to sea have been considered, however successful treatment has not proven to be cost beneficial to date. This study shows that sono-electrochemical treatment (combined electrolysis and power ultrasound) to produce magnesium hydroxide can raise the pH of the water, precipitate iron as insoluble iron hydroxide [Fe(OH)<sub>2</sub>] and has the potential to preferentially precipitate other metals in their stable hydroxide forms. Extrapolating the laboratory results and methods to full scale treatment (12 l sec<sup>-1</sup> flow rate) indicates that it is a viable Active treatment process compared to other treatment options being considered and can aid failing water bodies achieve compliance with the EU Water Framework Directive.

#### Introduction

Parys Mountain, N. Wales is one of the most polluting mine waters in the UK, discharging more metals into the Irish Sea than the River Mersey despite having less than 0.3% of the flow of the River Mersey. Parys has a number of discrete discharges of which Dyffryn Adda is considered the most challenging water chemistry being more acidic and metal rich than most surface water discharges in the UK. The Dyffryn Adda pollutes 3 km of the Afon Goch Amlwch before discharging into the sea at Port Offeiriad causing ochreous staining and depositing copper (c. 10 tonnes annum<sup>-1</sup>), zinc (c. 24 tonnes annum<sup>-1</sup>) and cadmium (~45 kg annum<sup>-1</sup>) into the Irish Sea. The discharge chemistry also includes other contaminants (Table 1).

Total (ugl <sup>-1</sup> )	Minimum	Mean	Maximum
Acidity (as CaCO <sub>3</sub> )	1,194	1,743	2,210
рН	2	3	3
Iron	453,000	599,000	708,000
Copper	30,600	38,130	52,600
Zinc	55,900	65,311	92,900
Manganese	14,300	20,470	30,400
Aluminium	4	63,571	87,800
Sulphate (as SO4)	1,940,000	2,534,000	3,020,000
Cadmium	153	175	196
Arsenic	129	360	662
Nickel	154	189	240
Lead	19	28	46

Table 1. Dyffryn Adda - Flow Concentrations (12 lsec-1 [1040 m3day-1])

A 'Parys Mountain Treatment Options Report' was commissioned in 2012 by Natural Resources Wales (formerly Environment Agency Wales) (URS 2012) to review treatment options to ameliorate the environmental impact of Parys Mountain on the economic and well being benefits to local and national communities. In further discussions it was considered that the first stage of amelioration would be to reduce the total iron discharged from Dyffryn Adda to <1.0 mgl<sup>-1</sup> and to reduce the ochreous visual impact

discharge in the Afon Goch Amlwch and into the sea, now the North Anglesey Marine possible Special Area of Conservation.

The releases of iron, sulphate and high levels of acidity are primarily a result of pyrite oxidation catalyzed by sulphur oxidizing bacteria (Barbes *et al.*, 1968):

 $4FeS_2 + 15O_2 + 14H_2O = 4Fe(OH)_3 + 8H_2SO_4$  (Sulphuric acid)

Traditional treatment methods for Mine water can generally be divided into 2 techniques:-

- 1. Active treatment which require ongoing mechanical / electrical operations and manual maintenance e.g. aeration, liquid chemical pH neutralization, chemical cationic coagulants and precipitation; membrane processes; ion exchange and chemical and / or biological sulphate removal;
- 2. Passive treatment typically refers to processes that do not require human intervention, operation, or maintenance, use gravity flow for water movement and promote the growth of natural vegetation e.g. reed bed systems.

Given the high concentrations of metals and low pH passive treatment at Parys Mountain is broadly accepted as being impractical and technically unfeasible. Active treatment comprising of High Density Sludge (HDS) with pH adjustment using lime is however widely used for treating acid mine drainage (AMD) water. In addition some success has been achieved using sulphide reduction (addition of sodium sulphide) after pH adjustment (caustic soda dosing) as a potential treatment option. In some circumstances hybrid active/passive systems may be employed e.g. semi-passive iron removal with targeted removal of other metals by sulphide precipitation or biological removal of iron as schwertmannite and removal of targeted metals by sulphide reduction along with part treatment and part pumped sea dispersal options. Budget costs for these various options are presented (Table 2).

Table 2. Cost Estimates for Iron removal / reduction from Dyffryn Adda adit

Element / Description	Cost Estimate	
Pumped sea dispersal of partially treated mine water	£ 0.6 to £ 0.7 M	
High Density Sludge (HDS) Plant	£ 1.8 to £ 2.0 M Capex / £ 0.5 to £ 0.7 M Opex	
Hybrid Active / Passive Treatment (incl. Sulphide treatment)	£ 1.8 to £ 2.5 M Capex / £ 0.25 M Opex	

The presence of metal species either in their ionic form or in equilibrium as their stable oxide, hydroxide forms, in a water phase such as mine water, are determined by their respective Pourbaix solubility diagrams (Pourbaix 1964). For example, the Pourbaix diagram for Iron in Water (Fig. 1) indicates the predominance of iron as aqueous ions ( $Fe^{2+}$ ,  $Fe^{3+}$ ) or as stable insoluble oxide or hydroxide forms [ $Fe(OH)_2$ ,  $Fe(OH)_3$ ] based on the pH of the water (horizontal axis) and its  $E_H$  (voltage potential) i.e. oxidizing or reducing potential (vertical axis).

For example in Dyffryn Adda (pH = 3.13 and  $E_H = 0.287$ ) iron will be present as soluble  $Fe^{2+}$  aqueous ions. If the mine water is aerated ( $E_H > 0.0$ ) and pH corrected to < pH 6.5 iron will precipitate as insoluble orange/brown ferric hydroxide [Fe(OH)<sub>3</sub>]. Under an oxygen reduced state ( $E_H < 0.0$ ) the iron will precipitate as green ferrous hydroxide [Fe(OH)<sub>2</sub>]. Similar reactions can be achieved within an electrochemical treatment process by selecting appropriate electrode materials and applying a voltage across the anode and cathode electrodes to control current density (Amps/Electrode area [Am<sup>-2</sup>]) (Fig. 2). Increasing – decreasing current density can be regarded as way of speeding up or slowing down reactions and depending upon the electrode material making reactions more oxidizing or more reducing.

By applying a voltage to an oxidizing over-potential anode electrode (e.g. MMO – mixed metal oxide) the water in the electrochemical reactor can be oxidized and made acidic.

 $2H_2O = H_2O_2$  (oxidized) +  $2H^+$  (acid) +  $2e^-$ 

Similarly by employing a voltage to an alkaline earth metal anode in water, the water becomes reduced and alkaline due to the formation of the hydroxide Mg(OH)<sub>2</sub>.

 $Mg^{2+}$  (anode) + 20H<sup>-</sup> (cathode) =  $Mg(OH)_2$ 





Fig 1. Iron – Water Pourbaix diagram

**Fig 2.** pH and E<sub>H</sub> variations within an electrochemical reactor

Once magnesium hydroxide is generated it operates by the principal of ion exchange, where a magnesium ion  $(Mg^{2+})$  exchanges with a metal ion  $(M^{2+})$  as shown by the following equation:

$$Mg(OH)_{2 (adsorbent)} + M^{2+} (aq.) = M(OH)_{2 (adsorbent)} + Mg^{2+}$$

Although the metal is shown to be divalent (<sup>2+</sup>) in this equation for simplicity, it may be of any valency provided that ion exists as a cation in solution. The magnesium ions ( $Mg^{2+}$ ) may further react with sulphate ions ( $SO_4^{2-}$ ) to produce insoluble magnesium sulphate.

$$Mg^{2+} + SO_4^{2-} = MgSO_4$$

In general, the tendency is for the solubility of the hydroxides of other treated metals to be lower than that of magnesium hydroxide as shown in Table 3 and to precipitate in preference to magnesium hydroxide.

Table 3. Solubility Data for Metals

Metal Hydroxide	KsP	Solubility (molel <sup>-1</sup> )
Copper - Cu(OH)2	2.2 x 10 <sup>-20</sup>	2.8 x 10 <sup>-7</sup>
Cadmium -Cd(OH)2	1.7 x 10 <sup>-15</sup>	1.2 x 10 <sup>-5</sup>
Chromium -Cr(OH)₃	1.7 x 10 <sup>-24</sup>	1.2 x 10 <sup>-8</sup>
Nickel - Ni(OH)2	6.5 x 10 <sup>-18</sup>	1.9 x 10 <sup>-6</sup>
Zinc - Zn(OH) <sub>2</sub>	1.7 x 10 <sup>-16</sup>	5.5 x 10 <sup>-6</sup>
Lead - Pb(OH)2	1.1 x 10 <sup>-20</sup>	2.2 x 10 <sup>-7</sup>
Iron - Fe(OH)2	8.8 x10 <sup>-16</sup>	6.0 x 10 <sup>-6</sup>
Magnesium - Mg(OH) <sub>2</sub>	1.1 x10 <sup>-11</sup>	2.2 x 10 <sup>-4</sup>

Traditional chemical methods of treating AMD using magnesium hydroxide are already described (Bologo *et al.*, 2009, 2012) and report good removal rates for metals from similar AMD waters in Witwatersrand Basin in S. Africa and other surface finishing and wastewater streams (Walter *et al.*, 2015). The sacrificial dissolution of a magnesium electrode by electrolysis to electro-generate magnesium hydroxide in-situ in a treatment process however has not been reported. The use of electrolysis, which encompasses electrocoagulation, -flocculation and -flotation, have previously been trialled and reported for treatment of AMD waters (Florence 2013) using iron, aluminium sacrificial electrodes and oxygen over-potential electrodes such as MMO and Ebonex (Hayfield 2001).

One major draw back with electrolysis (electro-coagulation) is passivation of the anode and cathode electrode surfaces during operation. Such fouling can lead to deterioration of treatment performance and excessive electrical voltages being used to achieve acceptable treatment current densities in the treatment process. Passivation is partially overcome in some electrochemical equipment by using high shear velocities across the electrode surfaces, polarity reversal and off-line electrode acid washing. This paper reports the novel use of sono-electrochemistry, electro-coagulation with combined power ultrasound (Morgan 2014) for treatment of AMD using a magnesium anode to produce magnesium hydroxide in-situ. Using power ultrasound simultaneously with electrolysis removes the ionic boundary and passivation layers (Stern and

Helmholtz layers) that can develop along the electrode surface during operation, making 'fresh' electrode material available for treatment. This reduces the electrical resistance of reactor circuit, reduces the power requirement and increases treatment efficiency and effectiveness.

# **Materials and Methods**

Water samples were collected from the sampling flume at Dyffryn Adda (GIS SH4380791223). Two litre aliquot samples were treated using a Soneco<sup>®</sup> (Power & Water) sono-electrochemical reactor consisting of stainless steel cathode, four sets of 28 kHz ultrasonic transducers, magnesium anode, two litre B-Ker square flocculator jar and Watson Marlow peristaltic recirculation pump. Water samples were re-circulated between the B-Ker and Soneco® reactor using the peristaltic pump. Following sonication and electrocoagulation, samples were flocculated on a Phipps & Bird bench flocculator for one minute at 250 rpm with addition of 5g of dried micro-sand ballast (<150 um diameter, stirred at 150 rpm with addition of 0.2ml polymer for three minutes and settled for 3 minutes at 0 rpm. 50ml aliquot samples were then filtered through a Whatman No. 10 filter before being tested for iron on a Hach-Lange DR 3900 spectrophotometer using the Hach Lange Iron Test kit (LCK 320). Due to limited time constraints and test methods the only other metal tested was copper using Quantofix Copper (0 – 100 mgl<sup>-1</sup>) dip-test.

Reaction time, pH (Hanna Instruments),  $E_{\rm H}$  (mV) (Hanna Instruments) and amperages were noted during the treatment procedure and photographs of the treatment reactions were taken. These were used along with the mechanical and electrical specifications of the Soneco® reactor to calculate Capex and Opex for a full-scale treatment plant.

## Results





Rapid

settlement

**Clarified** sample





Addition of

micro-sand and

polymer

400

300 200

0

-200

-300

(mV) 100

H -100

Table 6. Changes in sample through treatment process

Precipitation at

pH c. 3.5-3.7

Table 7. Iron reduction

Untreated

sample

Test	Amperage	Time	[Fe] <sub>Start</sub> (mgl <sup>-1</sup> )	[Fe] <sub>End</sub> (mgl <sup>-1</sup> )
Run 1	2	13	800	0.069
Run 2	1.5	17	800	0.007

Precipitation at

pH c. 7.5

Table 8. Copper reduction

Test	[Cu] Before mgl <sup>-1</sup>	[Cu] after mgl <sup>-1</sup>
Run 1	40	<1.0
Run 2	40	<1.0

Based on the above tests its is estimated that the magnesium  $(Mg^{2+})$  dissolution required to both pH neutralize and precipitate iron to <1 mgl<sup>-1</sup> is 325 mgl<sup>-1</sup>. Based on this ratio the predictive scale-up costs (capital and operating) based on a flow of 12 lsec<sup>-1</sup> will:-

- 1. Capex £840,000 GBP Soneco® Reactor & Power Supply excl. solid-liquid separation tank
- 2. Opex £220,000 GBP per annum Sono-electrochemical plant excl. replacement anodes

#### Discussion

The results confirm that electrical dissolution of a magnesium electrode produces magnesium hydroxide which exhibits the same characteristics as Mg(OH)<sub>2</sub> powders and granules currently employed in AMD treatment but overcomes the potential Health Safety and Environmental impacts of having liquid chemicals and pH correcting chemicals on site. The reaction raises the pH of acid waters, precipitates iron as ferrous hydroxide and other metal hydroxides. In common with Bologo *et. al.*, 2012 ferric hydroxide is first precipitated as a result of raising the pH from c. 3.0 pH units to 3.5-3.7 pH units to the insolubility product of Fe(OH)<sub>3</sub>. As the pH further rises and the water becomes reduced iron is finally precipitated as Fe(OH)<sub>2</sub>. The removal of copper during the reaction supports the findings that magnesium hydroxide will precipitate at 1 pH unit below their normal insolubility product. Accordingly it is likely that this treatment method should provide a 'sweep' removal of most alkaline mine water metals dissolved in Dyffryn Adda adit including cadmium.

Extrapolating the bench scale operating parameters and results to a full-scale treatment plant (12 lsec<sup>-1</sup> flow rate) indicates that the Capex and Opex will be £840,000 GBP and £220,000 GBP respectively. These costs are slightly lower than the budget costs already presented (Table 2, above).



Fig. 3 P&ID Sono-electrochemical treatment plant for AMD



Fig. 4 Soneco<sup>®</sup> Sono-electrochemical water treatment for removal of Phosphorus from wastewater

For future consideration, the Capex and Opex costs could be off-set by the recovery of metal from the sludge after treatment, and as reported by Bologo *et al.*, 2009, even the soluble magnesium in the treated water could be recovered by precipitating with carbon dioxide gas and recycled as magnesium hydroxide back to the head of works within the treatment process.

Further, the potential use of renewable energy sources at site could be used as an electrical power source for a sono-electrochemical treatment process.

For illustration purposes, a P&ID Sono-electrochemical treatment to treat Dyffryn Adda adit water is shown in (Fig. 3), together with a typical Soneco<sup>®</sup> (sono-electrochemical treatment plant) for removal of Phosphorus from municipal wastewater (Fig. 4). **Conclusion** 

- 1. Electro-generation of magnesium hydroxide by electrolysis of a magnesium electrode has been successfully demonstrated and that its properties for water treatment are similar to proprietary powders and granular media. These properties include:
  - a. Raising the pH of an acid water and use on AMD schemes;
  - b. Electro-generated magnesium hydroxide successfully precipitates Fe<sup>2+</sup> aqueous as insoluble ferrous hydroxide; the residual iron concentrations of the treated water being <1.0 mgl<sup>-1</sup>;
  - c. Magnesium hydroxide will simultaneously precipitate other metals as their insoluble hydroxide forms.
- 2. The predicted Capex and Opex of a sono-electrochemical treatment plant is broadly attractive to other treatment / disposal options being considered for Dyffryn Adda adit water. These costs could be off-set by the revenue earned form metals recovered from the treated sludge. Electrical energy for a sono-electrochemical plant could be sourced from a renewable energy supplies.
- 3. This present study was undertaken under a time constraint and as 'proof of concept' to quickly assess the viability of sono-electrochemistry (electro-generation of magnesium hydroxide) for water treatment for AMD. Having achieved favourable results and conclusions, further bench trials, extensive sampling testing and site pilot study would be recommended.
- 4. This technique is transferable and can be accordingly scales for metal mine water treatment projects at other mines challenged by steep terrain and exhibiting acid or ochreous mine discharges, like those at Cwm Rheidol and Cwmystwyth mines also being considered by Natural Resources Wales.

# References

URS (2012). Parys Mountain – Treatment Options Report. Draft 30 March 2012., Environment Agency Wales. Project Reference 46399817/MARP002.

Barbes, H.J. and Romberger, S.B. (1968). Chemical Aspects of Acid Mine Drainage. J. Water Poll. Contr. Fed. **40** (3) 371 – 384.

Pourbaix, M. (1964). Atlas of Electrochemical Equilibrium in Aqueous Solutions. Published by National Association of Corrosion (1974). ISBN 10: 0915567989.

Bologo, V., Maree, J.P. and Zvinowanda, C.M. (2009). Treatment of Acid Mine Drainage Using Magnesium Hydroxide. Abstracts of the International Mine Water Conference, Proceedings ISBN No. 878-0-9802623-5-3, 19<sup>th</sup>-23<sup>rd</sup> October 2009, Pretoria, South Africa.

Bologo, V., Maree, J.P., and Carlsson, F. (2012). Application of Magnesium Hydroxide and Barium Hydroxide for the removal of Metals and Sulphate from Mine Water. Water SA vol. **38** n.1 Pretoria.

Walter, M.D., Witkowski, J.T., and Gibson, A. (2015). Removal of Metals from Metal Finishing Waste Water Using a Granular, Magnesium-Based Adsorbent. Martin Marietta Magnesia Specialties, LLC. 8140 Corporate Drive, Suite 220, Baltimore, Maryland 21236 USA.

Florence, K.M, and Morgan, P.G. (2013). A novel electrochemical process for aqueous oxidation of acid mine drainage using advanced power electronics and real time control. IMWA Conference "Reliable Mine Water Technology" IMWA 31<sup>st</sup> May 2013, Golden CO; USA.

Hayfield, P.C.S. (2001). Development of a New Monolithic Ti4O7 Ebonex Ceramic. Published by Royal Society of Chemistry. ISBN 10: 0854049844.

Morgan, P.G. (2014). British Patent No. GB 1503638.7 – Method and apparatus for decontamination of fluids (Soneco). International Patent PCT/GB2016/050692 – Method and apparatus for decontamination of fluids.